# PHOTOCATALYTIC DEGRADATION OF ALIZARIN RED DYE IN AQUEOUS SOLUTION USING TITANIA-NICKEL AND TITANIA-COBALT NANOCOMPOSITES

## Odeyemi, O. T.<sup>a</sup>, Owalude, S. O.<sup>b</sup> and Odebunmi, E. O.\*<sup>b</sup>

<sup>a</sup>Department of Science Laboratory Technology, Kwara State Polytechnic, Ilorin, Nigeria <sup>b</sup>Department of Chemistry, University of Ilorin, P.M.B. 1515, Ilorin, Nigeria \*Corresponding Author e-mail address: sowalude@gmail.com (Received: 15th December, 2017; Accepted: 29th September, 2018)

#### **ABSTRACT**

Titanium dioxide (TiO<sub>2</sub>) was prepared by Sol-gel method which involved hydrolysis of Titanium trichloride (TiCl<sub>3</sub>) with concentrated hydrochloric acid and oxidation by hydrogen peroxide. Doping of the TiO<sub>2</sub> was effected at 10% Ni-TiO<sub>2</sub> and 10% Co-TiO<sub>2</sub> using aqueous solutions of nickel nitrate and cobalt nitrate respectively. Characterization by X-ray Diffraction Crystallography (XRD) and Scanning Electron Microscopy (SEM) showed that the surface area and crystallinity of TiO<sub>2</sub> increased through doping. The XRD and SEM data confirmed the observed increase in the photocatalytic activities of the doped versus the undoped TiO<sub>2</sub> photocatalysts. Photocatalytic activities of both the doped and undoped TiO<sub>2</sub>was investigated on the photodegradation of Alizarin red dye. With a constant dye concentration of 25 ppm and temperature of 40°C, it was observed that for all varying concentrations except 0.25 g/l of catalyst at 240 mins, the percentage degradation was of the order TiO<sub>2</sub>< Co-TiO<sub>2</sub>< Ni-TiO<sub>2</sub>. At the 0.25 g/l concentration of catalyst at 240 mins, the photocatalytic activities of the Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub> were observed respectively to be about 9 times and 12 times that of the TiO<sub>2</sub>.

Key Words: Titanium dioxide, titania-nickel, titania-cobalt, alizarin red and photocatalysts.

## **INTRODUCTION**

Water pollution caused by indiscriminate discharge of wastes into water bodies render the water unfit for human or biological use (Mann et al., 2014). Among these contaminants, organic dyes generated from dye-producing or dyeconsuming industries are frequently discharged into the water bodies. Dyes generally have complex aromatic molecular structures making them stable and extremely resistant towards biological or chemical degradation processes (Tehrani-Bagha et al., 2010; Nguyen and Juang 2013). Dyes as major organic water pollutants are classified into different groups according to their chemical components termed "chromophore". Therefore, degradation of any dyes are better studied according to the chromophore rather than other classes (Hunger et al., 2004).

Alizarin red (1,2-dihydroxyanthraquinone) is a type of dye having anthraquinone chromophore. It can be obtained naturally from the root of madder and synthetically from anthracene and used to make other forms of the dye e.g. Turkey red, Alizarin blue and Alizarin orange (Bien *et al.* 2005). A basic characteristic is its slight solubility in water; therefore in ancient times, it was used as a

histological stain. Alizarin red dye may result in potential risks of bioaccumulation, retardation in growth of aquatic organisms, adverse effect on photosynthesis by marine flora and cause threats to human health through the food chain transport (Hao *et al.*, 2000; Tayade *et al.* 2007).

Degradation of organic dye pollutants can be carried out using conventional physical, chemical and biological methods (Ashfaq and Khatoon, 2014). Some of these methods are limited by their high cost, inability to remove soluble dyes and addition of chemicals (Poyatos et al., 2010). Advanced oxidation processes have been discovered to completely degrade dye molecules from waste water. Of all these processes, photocatalysis emerged as the most promising treatment method for various dyes on a laboratory scale (Kulkarni and Thakur, 2014). Photocatalysis is a method that uses light to generate active species on light-sensitive compounds (photocatalysts) to effect water purification. Some of these active species are hydroxyl radicals (OH) which are generated in sufficient quantities to effect the degradation (Hao et al., 2000; Catrinescu et al., 2004; Eslami et al., 2007).

Literature reports showed that the activities of photocatalysts had been increased through doping (Rauf and Ashraf, 2009). Photocatalyst's doping involves the introduction of two or more elements (metals or non-metals) into the structure of catalyst which then results in the insertion of energy levels between the conduction and valence bands of the catalyst (Barolo *et al.* 2012). The inserted energy levels provide sub-band gap irradiation from which electrons can be excited from dopant d-band to conduction band or from valence band to dopant d-band by lower energy photons than are required by the pure photocatalysts (Ibhadon and Fitzpatrick, 2013).

In this work, Alizarin red dye was photodegraded by synthesized TiO<sub>2</sub>, Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub> photocatalysts. These photocatalysts were characterized by various analytical techniques and their photocatalytic activities for the degradation of Alizarin red dye were compared.

#### **MATERIALS AND METHODS**

Materials

Reagents used were Titanium trichloride (Tianjin Damao Chemicals, China), concentrated hydrochloric acid (BDH Laboratory Poole, England), 95% Ethanol (BDH Laboratory Poole, England), Ammonia solution and Hydrogen peroxide (Sigma-Aldrich). All reagents were of analytical grade and were used without further purification.

## Synthesis of doped and undoped TiO<sub>2</sub>

The experimental method used was the Sol-gel method described by Molea and Propescu (2011) and Wang et al. (2015). Distilled water and 10% TiCl<sub>3</sub> containing 15% HCl solution (50 ml each) were added together with 75 ml 30% NH<sub>3</sub> solution in a 500 ml flat-bottomed flask. The solution was continuously stirred after which it was refluxed for 24 h at 70°C. The resultant solution was allowed to cool, thereafter 75 ml of H<sub>2</sub>O<sub>2</sub> was added. A yellow precipitate developed which was aged for 12 hours to form titaniumperoxo complex solution. The solution was then heated for 4hours at 70°C to give a gel-like substance which was dissolved in water, autoclaved for 24 hours and then oven-dried for 1hour at 70°C. The dried product was calcined at 450°C for 1hour and the titanium dioxide powder obtained was stored in air tight bottles.

The doped TiO<sub>2</sub> catalysts were prepared at concentration of 10% each for the nickel and cobalt as dopants. To prepare 10% Ni-TiO<sub>2</sub>, 0.50 g of Ni(NO<sub>3</sub>)<sub>2</sub> salt was added to 100 ml of a mixture of distilled water and ethanol (3:1v/v) (Kumar *et al.*, 2015) and 4.50 g of the TiO<sub>2</sub> was added to the metal solution and stirred. The solution was then refluxed for 4 hours at 70°C. The resulting solution was cooled and oven-dried at 70°C for 4hoursand then calcined for 1hour. The nickel-doped titanium dioxide powder was stored in an air tight bottle. The process was repeated for 10% Co-TiO<sub>2</sub> using its corresponding nitrate salt.

### Characterization

The synthesized photocatalysts were characterized using Scanning Electron Microscopy and X-ray Diffraction methods. The Scanning Electron Microscopy (SEM) was recorded on a Scanning Electron microscope (Aspex 3020) at an accelerating voltage of 16.0 kV and the images were obtained at a magnification of 250x. The X-ray diffraction (XRD) patterns were recorded on a powder X-ray diffractometer (Bruker D8 Advance) using Cu-K\(\alpha\) radiation. Diffraction patterns were recorded in the 2\(\theta\) scale ranging from 10° to 90° at step interval of 0.034°.

#### Photocatalytic activities

A 25 ppm solution of Alizarin red (AR) dye was prepared using distilled water. The dye degradation began with varying concentrations of the photocatalysts ranging from 0.10 g/l to 0.30 g/l at a constant temperature of 40°C. The solution was transferred to the reactor which was irradiated with a UV lamp. About 5 cm<sup>3</sup> sample was withdrawn at regular time intervals of 30 mins up to a maximum of 240 mins. The Absorbance of each sample was measured at 340nm, the wavelength corresponding to maximum absorbance. The experiment was repeated for the doped photocatalysts. The absorbance of the dye solution at 40°C when no catalyst was present was also measured as blank, and the percentage degradation was calculated at each sampling time using the equation:

% degradation = 
$$\frac{A_o - A_t}{A_o} X$$
 100

where  $A_{\circ}$  and  $A_{\circ}$  are the initial absorbance (blank absorbance) and absorbance after various intervals of time, t respectively. The photocatalytic activities of the doped and undoped photocatalysts were thus measured and compared.

#### RESULTS AND DISCUSSION

The Scanning Electron Microscopy (SEM) was used to observe the morphological changes that occurred as a result of doping of the TiO2 photocatalysts by the nickel and cobalt metals. The SEM micrographs are presented in Fig. 1.It was observed that all the catalysts were similar in shape. However, the dopedTiO2had smaller particles than the undoped. TiO, can be said to slightly fragment at the incorporation of the nickel and cobalt metals which led to increase in surface area of the doped catalysts. With increased surface area of the catalyst, increased degradation is expected on the part of the doped catalysts (Chanderia et al., 2017). The particle sizes of the TiO<sub>2</sub>, Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub> are 16.6 mm, 16.5 mm and 16.0 mm respectively.

The XRD patterns of all the catalysts are presented in Fig. 2. The XRD patterns of Ni-TiO<sub>2</sub> and Co-TiO2 exhibited strong diffraction peaks at 25.2°, 37.8°, 48.1°, 53.8° and 62.6° confirming anatase TiO2 while the XRD pattern of the undoped TiO<sub>2</sub> exhibited strong diffraction peaks at 27.4°, 36.1°, 41.2°, 54.3° and 69.0° confirming rutile TiO, (Munusamy et al., 2013). It was observed that the peak intensity of diffraction pattern increased proportionally then decreased from undoped to the doped. Therefore, doping with both Nickel and Cobalt metals altered the crystal form of TiO<sub>2</sub> from rutile to anatase phase. The results also showed that the crystal plane (110) for TiO<sub>2</sub> appeared at 27.5° while the crystal plane (101) for both Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub> appeared at 25.3° which are all in agreement with literature reports and spectrum of the JCPDS (Joint Committee on Powder Diffraction Standards)(Li and Chen, 2011; Landmann et al., 2012). The crystallites sizes of all the photocatalysts were calculated using the Scherrer equation and were observed to be 0.9 nm, 1.41 nm and 1.41 nm respectively for the TiO<sub>2</sub>, Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub>.

The photocatalytic activities of the samples were distinguished based on the catalyst loading. These are presented in Fig. 3 and summarized by the data presented in Table 1. The amount of catalyst used for the dye degradation was effected in the range 0.10 g/l - 0.30 g/l. The % degradation of the undoped and doped TiO2 were observed to increase as the amount of catalyst increased up to 0.25 g/l. Literature report indicate that the surface area of catalyst available for degradation increased with increase in the amount of the catalyst leading to a rise in the rate of reaction is usually observed (Chanderia et al., 2017). The increase in the rate of dye degradation has also been attributed to the availability of more sites on the catalyst surface until an optimum level was reached (Munusamy et al., 2013). It was observed from Table 1 that the rate of dye degradation tends to decrease at high catalysts' loading. This decrease has been attributed to the agglomeration of catalyst particles at higher concentration than the optimum level (Munusamy et al. 2013). The decrease in % degradation at high catalyst loading has also been attributed to the reduction in the penetration of light with increase in the amount of catalyst beyond the critical point (Mahadwad et al. 2011).

Catalyst loading of 0.25 g/l of undoped TiO<sub>2</sub> presented about 20% degradation at 60 mins on stream. This % degradation decrease to about 6 times its initial % degradation as the time increased to 240 mins. It shows that the maximum degradation achievable by the undoped catalyst was 20% at 60 mins. However, 0.25 g/l of the doped TiO<sub>2</sub> presented 21.6% in Ni-TiO<sub>2</sub> and 22.02% in Co-TiO2 at 60 mins and increased respectively to about 1.4 times and 1.9 times their initial degradation at 240 mins. This implies that the doped forms increased the catalytic activities of the undoped form beyond its degradable efficiency with time. This is attributed to the ability of dopants to suppress the recombination of photo-induced electron-hole pairs thereby increasing the photocatalytic efficiency in the TiO<sub>2</sub>(Zhang and Liu, 2008). In essence, it can be concluded that the photocatalytic activities of the Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub> were respectively about 9 times and 12 times that of the TiO<sub>2</sub> in the present study

It was also observed that Ni-TiO<sub>2</sub> showed a higher degradation of 35.51% at catalyst loading of 0.2 g/l than its expected optimum concentration of 0.25 g/l. This may imply that doping with respect to nickel metal shifted the expected optimum

concentration to a lower concentration of catalyst while an optimum degradation was maintained. Therefore, the 0.20 g/l Ni-TiO<sub>2</sub> possess equal characteristics as 0.25 g/l Co-TiO<sub>2</sub> and undoped TiO<sub>2</sub>.

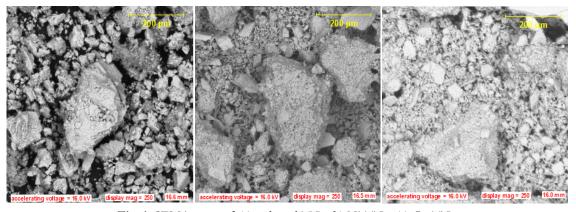


Fig. 1: SEM images of (a) undoped  ${\rm TiO_2}$  (b) Ni-TiO<sub>2</sub>, (c) Co-TiO<sub>2</sub>

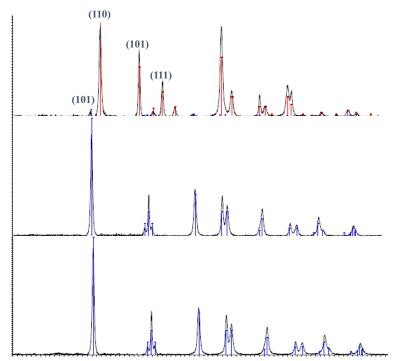


Fig. 2: XRD patterns of (a) undoped TiO<sub>2</sub> (b) Ni--TiO<sub>2</sub> (c) Co-TiO<sub>2</sub>

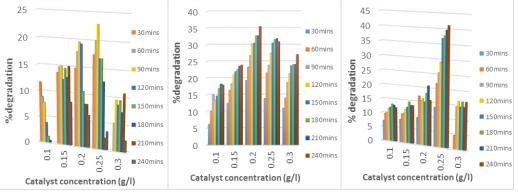


Fig. 3: Effect of catalyst concentration on the photocatalytic degradation of 25 ppm Alizarin red dye at 40°C using (a) Undoped TiO<sub>2</sub> (b) Ni-TiO<sub>2</sub> and (c) Co-TiO<sub>2</sub>

	% degradation											
[catalyst] (g/l) /time (mins)	TiO <sub>2</sub>				10% Ni-TiO <sub>2</sub>				10% Co-TiO <sub>2</sub>			
	60	120	180	240	60	120	180	240	60	120	180	240
0.3	5.24	8.73	7.42	2.18	14.05	21.41	24.08	27.09	4.8	16.61	16.61	16.61
0.25	20.06	16.88	12.74	3.5	21.6	27.58	31.56	30.9	22.02	29.28	37.92	41.28
0.2	17.9	19.28	8.25	6.19	23.37	26.79	32.71	35.51	17.11	16.35	18.25	15.97
0.15	14.83	12.5	12.92	8.33	16.39	20.98	22.62	23.93	10.49	11.97	14.67	13.51
0.1	8.92	3.88	0.39	-	10.1	13.36	16.94	17.92	9.77	11.72	13.28	12.11

Table 1: Summary of photocatalytic activities of TiO<sub>2</sub>, Ni-TiO<sub>2</sub> and Co-TiO<sub>2</sub>

#### **CONCLUSION**

TiO<sub>2</sub> has been shown to be an effective catalyst for the degradation of Alizarin red dye. The photocatalytic degradation of the TiO<sub>2</sub> was enhanced by doping. The % degradation of both the doped and undoped catalysts was observed to be influenced by the amount of catalyst at constant dye concentration and temperature.

#### REFERENCES

- Ashfaq, A. and Khatoon, A. 2014. Waste Management of Textiles: A Solution to the Environmental Pollution. *International Journal of Current Microbiology and Applied Sciences*. 3(7):780-787.
- Barolo, G., Livraghi, S., Chiesa, M., Paganini, M. C., Giamello, E. 2012. "Mechanism of the Photoactivity under visible light of N-doped Titanium dioxide: Charge carriers migration in irradiated N-TiO<sub>2</sub> investigated by electron paramagnetic resonance. *Journal of Physical Chemistry C*, 116:20887-20894.
- Bien, H.S, Stawitz, J. and Wunderlich K. 2005. Anthraquinone Dyes and Intermediates. Ullmann's Encyclopedia of Industrial Chemistry, Weinheim: Wile-VCH, p 355.
- Catrinescu, C., Cha S. Y. and Yang, H. Y. 2004. Improvement of oxidative decomposition of aqueous phenol by microwave irradiation in UV/H<sub>2</sub>O<sub>2</sub> process and kinetic study. *Water Research* 38:2782-2790.
- Chanderia, K., Kumar, S., Sharma, J., Ameta, R. and Punjabi, P.B. 2017. Degradation of sunset yellow FCF using copper-loaded bentonite and H<sub>2</sub>O<sub>2</sub> as photo-Fenton-like reagent." *Arabian Journal of Chemistry*, 10: s205-s211.

- Eslami, A., Nasseri, S., Yadollahi, B., Mesdaghinia, A., Vaezi, F. and Nabizadeh, R. 2007. Application of photocatalytic process for removal of methyl *tert*-butyl ether from highly contaminated water. *Iranian Journal of Environmental Health Science & Engineering* 4:215-222.
- Hao, O.J., Kim, H. and Chaing, P.C. 2000. Decolorization of wastewater. *Critical Reviews. Environmental Science and Technology* 30: 449-505.
- Hunger, K., Gregory, P., Miederer, P., Berneth, H., Heid, C. and Mennicke, W. 2004. Important chemical chromophores of dye classes in industrial dyes: Chemistry, properties, applications. (Ed. K. Hunger), Wiley-VCH Verlag, GmbH & Co. kGaA, Weinheim, FRG.
- Ibhadon, A. O. and Fitzpatrick, P. 2013. Heterogeneous photocatalysis: Recent advances and applications. *Catalysts* 3:189-218.
- Kulkarni, M. and Thakur, P. 2014. Photocatalytic degradation and mineralization of reactive textile azo dye using semiconductor metal oxide nanoparticles. *International Journal of Engineering Research and General Science* 2(2): 245-254.
- Kumar, A., Hitkari, G., Singh, S., Gautam, M.and Pandey, G. 2015. Synthesis of Ni-TiO<sub>2</sub>nanocomposites and photocatalytic degradation of oxalic acid in wastewater. *International Journal of Innovative Research in Science, Engineering and Technology* 4(12):12722-12731.
- Landmann, M., Rauls, E. and Schmidt, W.G. 2012.

  The electronic structure and optical response of rutile, anatase and brookite TiO<sub>2</sub>. Journal of Physics: Condensed Matter 24:1-

6.

- Li, Y. and Chen, W. 2011. Photocatalytic degradation of Rhodamine B using nanocrystalline TiO<sub>2</sub>-zeolite surface composite catalysts: effects of photocatalytic condition on degradation efficiency. *Catalysis Science & Technology* 1:802-809.
- Mahadwad, O. K., Parikh, P. A., Jasra, R. V. and Patil, C. 2011. Photocatalytic degradation of reactive black-5 dye using TiO<sub>2</sub> impregnated ZSM-5. *Bulletin of Materials Science* 34: 551-556.
- Mann, U.S., Dhingra, A. and Singh, J. 2014. Water pollution: causes, effects and remedies. International Journal of Advanced Technology in Engineering and Science 2: 1-5.
- Molea, A. and Popescu, V. 2011. The obtaining of titanium dioxide nanocrystalline powders. *Journal of Optoelectronics and Advanced Materials-Rapid Communications*5: 242-246.
- Munusamy, S., Aparna, R.S. and Prasad, R.G. 2013. Photocatalytic Effect of TiO<sub>2</sub> and the effect of Dopants on Degradation of Brilliant green. *Sustainable Chemical Processes*, 1:4: 1-8.
- Nguyen, T. A. and Juang, R. S. 2013. Treatment of waters and waste waters containing sulfur

- dyes A review. *Chemical Engineering Journal* 219: 109-117.
- Poyatos, J. M., Munio, M. M., Almecija, M. C., Torres, J. C., Hontoria, E. and Osorio, F. 2010. Advanced oxidation processes for wastewater treatment: State of the art." *Water, Air and Soil Pollution* 205:187-204.
- Rauf, M. A. and Ashraf, S. S. 2009. Fundamental principles and applications of heterogenous photocatalytic degradation of dyes in solution. *Chemical Engineering Journal* 151: 10-18.
- Tehrani-Bagha, A. R., Mahmoodi, N. M and Menger, F. M. 2010. Degradation of a persistent organic dye from colored textile wastewater by ozonation. *Desalination* 260: 34-38.
- Wang Q, Liu X, Wei, Dai J. and Li, W. 2015. Ferromagnetic Property of Co and Ni Doped TiO<sub>2</sub> Nanoparticles. *Journal of Nanomaterials* 2015: 1-5.
- Zhang, X. and Liu, Q. 2008. Visible-light-induced degradation of formaldehyde over titania photocatalyst Co-doped with nitrogen and nickel. *Applied Surface Science* 254: 4780–4785.